
Elemental and Radioactivity Concentration of Stream Sediments in Abu-Rusheid, Nugrus Area – South Eastern Desert, Egypt

S.U. El-kameesy¹, S.Y. Afifi², A. Hamid³, Ali Ajeeb^{1,*}

¹Department of Physics, Faculty of Science, Ain Shams University, Cairo, Egypt

²Nuclear Materials Authority (NMA), Cairo, Egypt

³Radiated Pollution Department, Hot Laboratories Center, Atomic Energy Authority, Egypt

Email address:

ali_ajeab@yahoo.com (A. Ajeeb)

To cite this article:

S.U. El-kameesy, S.Y. Afifi, A. Hamid, Ali Ajeeb. Elemental and Radioactivity Concentration of Stream Sediments in Abu-Rusheid, Nugrus Area – South Eastern Desert, Egypt. *American Journal of Physics and Applications*. Vol. 3, No. 6, 2015, pp. 183-189.

doi: 10.11648/j.ajpa.20150306.11

Abstract: As a part of the national survey to evaluate some strategic elements such as U, Mn, Fe and Cu we have undertaken a quantitative study to fulfill this target. Twenty sediment samples were collected from Abu Rusheid Area in South Eastern desert of Egypt. The concentrations of radium-226 and thorium-232 were determined by direct counting using a hyper pure germanium (HPGe) detector interfaced with a multi-channel analyzer (MCA). The obtained results showed that the specific activity of ²²⁶Ra ranged from 65.71±16 to 208.2±20 Bq/kg and ²³²Th from 27.14±8.69 to 206.89±40.1 Bq/kg while ⁴⁰K from 127.73 to 443.84 Bq/kg. These values are higher than the international recommended levels. The radium equivalent activity (Ra_{eq}), the absorbed dose rate (D), the external hazard index (Hex), the annual effective dose equivalent (AEDE) and the representative level index (I_{yr}) were also calculated and compared with the international recommended values. Additionally, X-ray fluorescence analysis for soil sample has been performed and considerable concentrations of Al, Fe, Mn and Zn has been found.

Keywords: ²²⁶Ra, ²³²Th, ⁴⁰K, Norm, South Eastern, Absorbed Dose Rates, Radium Equivalent Activity, External Hazard Index

1. Introduction

It is very well-known that every living creature is exposed to ionizing radiation. This radiation is a part of the earth and comes from cosmos and affects all the goods, food and even the air that we breathe and makes them partially radioactive. So, human beings are exposed to natural background radiation every day from the ground, building materials, air, food, the universe, and even elements in their own bodies [1].

The terrestrial component of the natural background is dependent on the compositions of the soils and rocks containing natural radionuclides. The radioactivity of soils is essential for understanding changes in the natural radiation background [2, 3]. Soil contains small quantities of radioactive elements Ra and Th along with their progenies. The natural radioactivity of soil and sediment depends on the soil and sediment formation and transport processes that were involved since soil and sediment formation; chemical and

biochemical interactions influence the distribution patterns of uranium, thorium and their decay products [4].

Abu Rusheid Area in South Eastern desert of Egypt is located between a major thrust to the NE and a minor one to the SW. The stream sediments along Nugrus area are formed due to the weathering processes of the different surrounding rocks. These rocks are almost of granitic origin. Granitic rocks have relatively high concentration of the members of the natural radioactive series of ²³⁸U, ²³²Th and ²³⁵U. The presence of these naturally occurring radionuclides along with the non-series radionuclide ⁴⁰K and their daughter products in building materials is a source of indoor radioactive pollution. Nugrus area is recently subjected to several geological activities to explore precious minerals such as gold, zirconium and pyrite, etc... This reflects the possibility of establishing many development projects at that area [5].

2. Materials and Methods

2.1. Sampling and Sample Preparation

Twenty representative geological samples were collected from Abu Rusheid, Nugrus Area in South Eastern desert of Egypt. The location of this area is clearly shown in Fig (1). All samples were dried at 105°C for 24 hours to eliminate

water of moisture. The representative powdered samples were placed in polyethylene bottles of 350 cm³. Each sample was sealed from air in a container, whose inner diameter was equal to the diameter of the detector in face to face geometry. Finally, samples were stored for four weeks to reach the equilibrium state between radium and its decay products.

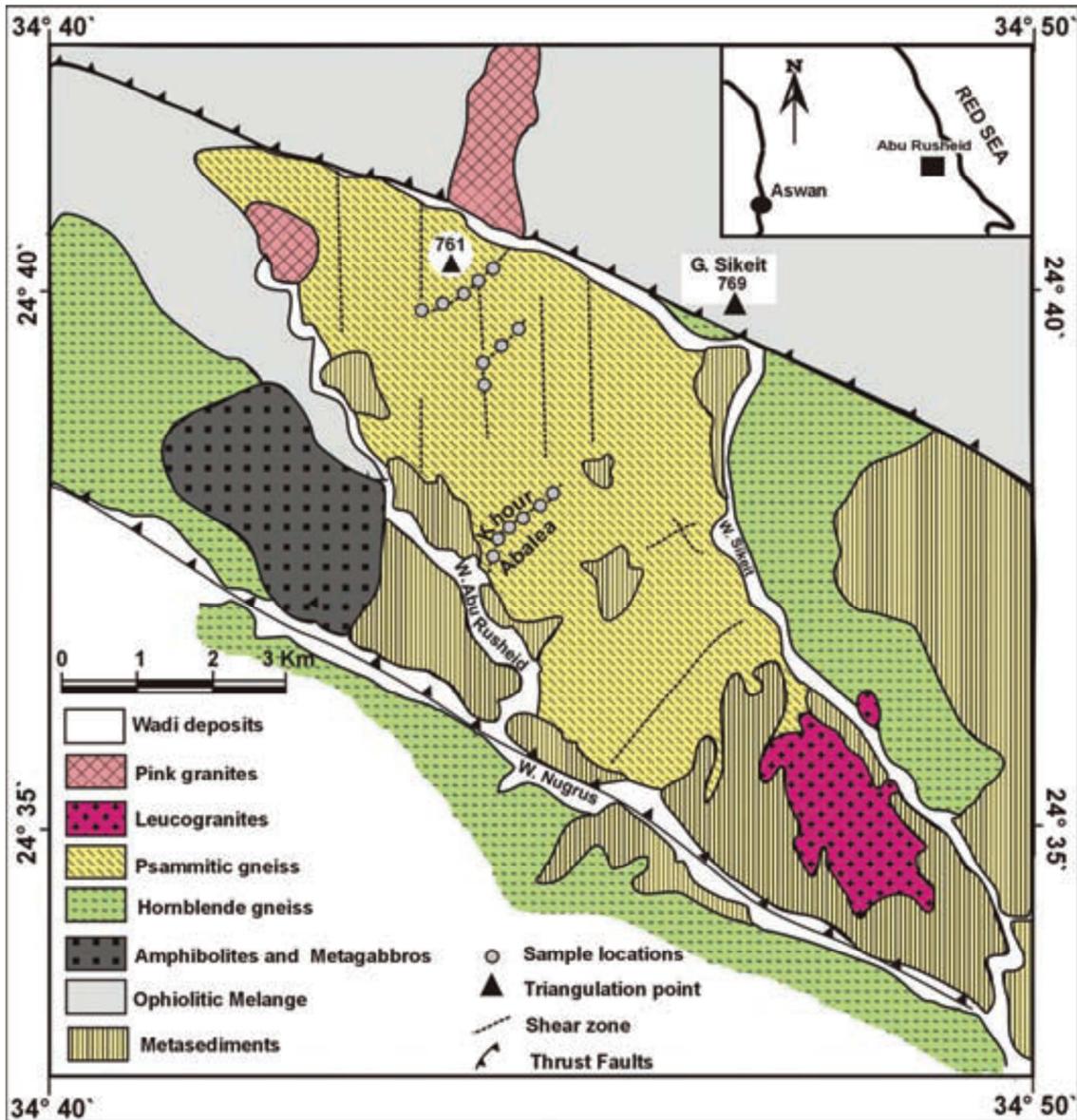


Figure 1. Geological map of the study area showing sample locations.

2.2. Experimental Techniques (Detection System Set-Up)

The energy and intensity of various gamma ray lines have been measured using an ORTEC HPGe detector of volume 62.3 cc and relative efficiency 30% relative to 3" x 3" NaI (Tl) detector coupled to a 4096 multichannel analyzer. The full width at half maximum (FWHM) was found to be 1.8 keV for ⁶⁰Co-1332 keV γ -ray line. For accurate energy determination, the spectrometer was calibrated using gamma ray lines of different standard sources. The activity concentration of ²²⁶Ra was determined using the 295.1

(19.2 %) and 352 (37.1 %) keV gamma-rays from ²¹⁴Pb and the 609.3 (46.1 %) and 1120.3 (15%) keV gamma-rays from ²¹⁴Bi. ²³²Th activity was determined from the gamma-rays, 238.6 (43.6 %) keV from ²¹²Pb, 338.4 (12%), 911.2 (29 %) and 969 (17%) keV from ²²⁸Ac and 583.0 (86 %), 2614 (36%) keV gamma-rays from ²⁰⁸Tl.

2.3. Description of the X-Ray Fluorescence System

XRF technology provides one of the simplest, most accurate and most economic analytical methods for the

determination of the chemical composition of many types of materials. It is non-destructive and reliable, requires no, or very little, sample preparation and is suitable for solid, liquid and powdered samples [6].

In this study, the XRF technique was used to determine the trace element contents using PHILIPS X'Unique-II spectrometer with automatic sample changer PW 1510, (30 positions). This instrument is connected to a computer system using X-40 program. The trace elements concentrations are calculated from the program's calibration curves which were set up according to international reference materials, (standards), as NIM-G, G-2, GSP-1, AGV-1, JB-1 and NIM-D [7].

The trace elements were measured by calibrating the system under the conditions of Rh-target tube, LiF-420 crystal, gas flow proportional counter, (GFPC), coarse collimators, vacuum, 30 kV and 40 mA for the determination of V, Cr, Co, Ni, Cu, Zn and Ga, 70 kV and 15 mA, for Rb, Sr, Y, Zr and Nb and 100 kV and 10 mA for the determination of Ba and Pb. The detection limit is the lowest concentration, and it is a function of the level of background noise relative to an element signal. The detection limit for the elements measured by XRF technique is estimated at 2 ppm for Rb, Nb, Ga, Co, Y and Sr and at 8 ppm for Pb and Cu and 5 ppm for other measured trace elements.

2.4. Sample Preparation for X-Ray Fluorescence

A total of 5 samples were completely analyzed using X-ray fluorescence, these samples were collected from Abu Rusheid, Nugrus Area in Eastern Desert of Egypt. For trace element analysis, pressed powder pellets were prepared by filling an alumina cup, (diameter 4 cm, height 1.2 cm and weight 3 gm), with 9 gm of crystalline boric acid covered by 1 gm of the grounded sample, (~200 mesh grain size), and then pressed under 12 t using semi-automatic hydraulic press model HERZOG HTP-40.

2.5. Calculation of Radioactivity Concentration

The specific activity (in Bq·kg⁻¹) is given by [8]:

$$A = \frac{NP}{t_c \cdot I_\gamma(E_\gamma) \cdot \epsilon(E_\gamma) \cdot M}$$

where NP is the net number of counts in a given peak area at energy E, $\epsilon(E_\gamma)$ the detection efficiency at energy E, t_c is the counting time in seconds, $I_\gamma(E_\gamma)$ the number of gammas per disintegration of this nuclide for a transition at energy E, and M is the mass in kg of the measured sample.

3. Results and Discussion

3.1. Radioactivity Concentrations

The activity concentrations of the radionuclides ²³²Th, ²²⁶Ra and ⁴⁰K in the collected samples, considered in the present study are presented in Table (1) and illustrated in Fig (2). The values are given in Bq/Kg on a dry weight basis. The

minimum and maximum along with the average values of the activity concentrations of the radionuclides are given in Table (2). The activity concentration of ²³²Th ranged from 27.14 to 206.89 Bq/kg with an average value 83.51 Bq/kg, which is higher than the international radioactivity average level (50 Bq/kg). The ²²⁶Ra activity concentration ranged from 65.71 to 208.23 Bq/kg with an average 120.23 Bq/kg, which is higher than the average values of the international recommended levels (35 Bq/kg) [9]. For ⁴⁰K, the average value is 299.082 Bq/kg, which is lower than the international average radioactivity levels (500 Bq/kg).

Fig 3 illustrates the correlation between the activity concentrations of ²²⁶Ra and ²³²Th. From that Figure It can be seen that there is a satisfactory correlation between ²²⁶Ra and ²³²Th with a correlation coefficient of 0.76. According to Tzortzis and Al-Hamarneh [10, 11], the theoretical values of the elemental ratios of Th/Ra are expected to be approximately 3.0 for normal continental crust. In the current study, the obtained results of the elemental ratios for Th/Ra varies from 0.413 to 1.104. The small value of the Th/Ra ratios may be attributed to a reduction of ²³²Th or an increase of ²²⁶Ra due to the alteration or natural processes in that area.

Table 1. The calculated specific activities (Bq/kg) for ²³²Th, ²²⁶Ra, and ⁴⁰K along with the ratio ²³²Th/²²⁶Ra in the investigated region.

Sample	²³² Th	²²⁶ Ra	⁴⁰ K	²³² Th/ ²²⁶ R
Sh-1	27.14±8.69	65.71±16	225.97	0.413
Sh-2	46.08±14.58	68.92±9.39	214.59	0.669
Sh-3	53.08±22.68	103.90±8.25	275.29	0.511
Sh-4	48.83±18.46	81.82±7.29	287.01	0.597
Sh-5	53.85±15.88	106.29±12.79	359.39	0.507
Sh-6	58.49±21.49	128.54±18.9	401.87	0.455
Sh-7	56.54±15.25	106.45±15.97	363.53	0.531
Sh-8	54.16±12.64	101.11±15.39	374.82	0.536
Sh-9	80.31±24.12	138.43±17.70	443.84	0.580
Sh-10	69.44±20.29	115.97±14.45	365.05	0.599
Sh-11	182.53±58.14	180.77±30.09	176.13	1.010
Sh-12	70.60±26.99	129.67±14.57	403.24	0.544
Sh-13	96.22±39.66	109.75±21.56	127.73	0.877
Sh-14	66.68±13.01	81.83±8.03	215.59	0.815
Sh-15	76.18±13.77	128.39±11.80	360.01	0.593
Sh-16	77.83±14.25	133.73±14.7	364.4	0.582
Sh-17	114.42±22.01	165.04±15.70	440.01	0.693
Sh-18	206.89±40.01	208.23±20.09	180.13	0.994
Sh-19	145.74±30.65	132.01±18.65	129.03	1.104
Sh-20	85.12±20.3	118.13±9.03	274.01	0.721

Table 2. Minimum, maximum and average activity concentration values for ²³²Th, ²²⁶Ra, ⁴⁰K and ²³²Th/²²⁶Ra.

nuclide	Minimum	maximum	Average
²³² Th	27.14±8.69	206.89±40.01	83.51
²²⁶ Ra	65.71±16	208.23±20.09	120.23
⁴⁰ K	127.73	443.84	299.082
Th/Ra	0.413	1.104	0.666

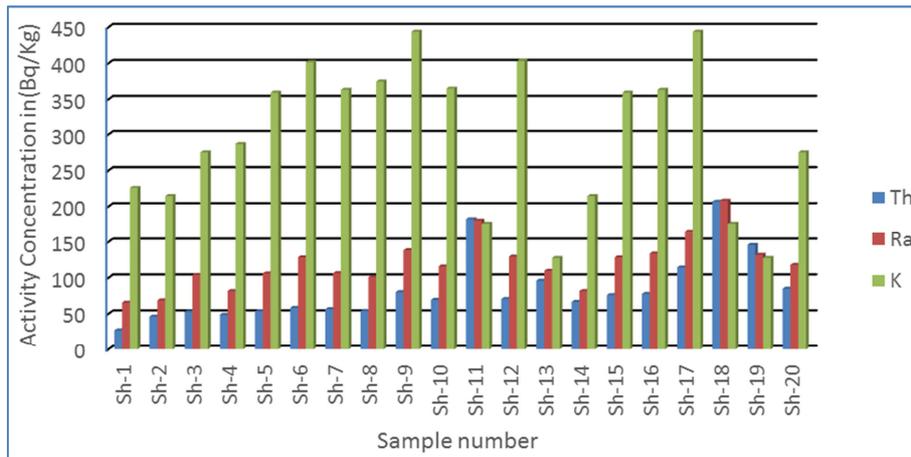


Figure 2. The activity concentration of radionuclides in the considered samples.

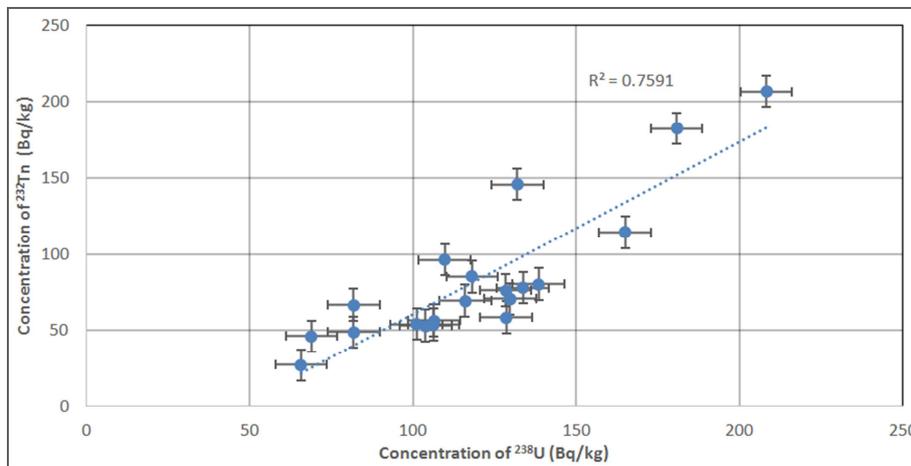


Figure 3. Correlation between ²²⁶Ra and ²³²Th activity concentration in all soil samples.

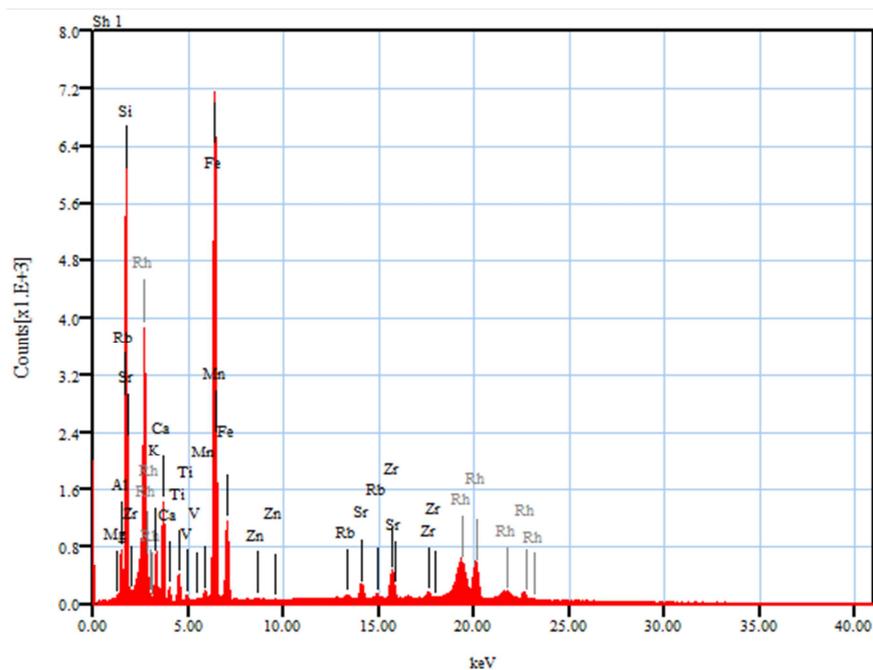


Figure 4. X-ray fluorescence (XRF) spectra of sample sh-1.

3.2. The Trace Element Analysis

According to the XRF results Mg, Al, Fe and Zn in some samples have remarkable concentration that could be extracted by different suitable methods in order to be used in different important industries. Mg contribution in the samples is ranging from 1.24 to 2.35 %, Al is ranging from 8.99 to 9.42 %, Fe is ranging from 12.84 to 18.76%, and Zn is ranging from 850 ppm to 1720 ppm. These values indicate that the rock samples of the place under study can be considered as a production area for the aforementioned elements. The analyzed samples are given in Table (3), and an example of the XRF spectra is displayed in Fig (4).

3.3. Assessment of Radiological Hazard

One of the main objectives of the radioactivity measurements in an environmental sample is not simply to determine the activity concentrations of ^{226}Ra , ^{232}Th and ^{40}K but also to estimate the radiation exposure dose and to assess the biological effects on humans. The assessment of radiological risk can be considered in various terms. In the current study, four related quantities were deduced, these are: (i) the absorbed dose rate (D) in air at 1 meter above the ground surface; (ii) the annual effective dose equivalent (AEDE) from outdoor terrestrial gamma radiation; (iii) the radium equivalent activity (R_{eq}); and (iv) the external hazard index (H_{ex}). These radiological parameters can be calculated from the measured activity concentrations of three main primordial radionuclides in the soil samples [12].

Table 3. Trace elements concentrations (percent), for the collected samples from Abu Rusheid area, Southeastern Desert of Egypt.

Element	Sh-1	Sh-4	Sh-8	Sh-9	Sh-10
Mg	1.24	1.61	1.68	2.35	1.412
AL	9.27	9.42	8.99	9.18	9.415
Si	51.356	54.08	57.01	56.43	57.18
K	6.27	9.29	9.8	9.21	9.35
Ca	10.23	8.9	7.32	8.09	6.97
TI	1.94	1.27	0.967	0.859	0.85
V	0.0248	0	0	0	
Mn	0.2114	0.203	0.18	0.167	0.126
Fe	18.76	14.29	13.2	12.84	13.78
Zn	0.085	0.136	0.11	0.172	0.129
Rb	0.062	0.145	0.148	0.173	0.145
Sr	0.187	0.103	0.09	0.08	0.079
Zr	0.341	0.281	0.228	0.272	0.322
Nb	0	0.052	0	0.095	0.056
Ru	0	0.153	0.086	0	
Cr	0	0	0.076	0.08	0.08
Ni	0	0	0.059	0	
Y	0	0	0.05	0	0.066
As	0	0	0	0	0.343

3.3.1. Absorbed Dose Rate in Air (D)

The absorbed dose rate could be obtained by the following formula [13]:

$$D \text{ (nGy.h}^{-1}\text{)} = 0.462A_{\text{Ra}} + 0.604A_{\text{Th}} + 0.0417A_{\text{K}}$$

Where D is the absorbed dose rate in nGy.h^{-1} , A_{Ra} , A_{Th} and A_{K} are the activity concentrations of ^{226}Ra , ^{232}Th and ^{40}K , respectively.

3.3.2. Annual Effective Dose Equivalent (AEDE)

The absorbed dose rate in air at 1 meter above the ground surface does not directly provide the radiological risk to which an individual is exposed. The absorbed dose can be considered in terms of the annual effective dose equivalent from outdoor terrestrial gamma radiation which is converted from the absorbed dose by taking into account two factors, namely the conversion coefficient from absorbed dose in air to effective dose and the outdoor occupancy factor. The annual effective dose equivalent can be estimated using the following formula [14]:

$$\text{AEDE } (\mu\text{Sv.y}^{-1}) = D \text{ (nGy.h}^{-1}\text{)} \times 8760 \text{ h} \times 0.2 \times 0.7 \text{ Sv.Gy}^{-1} \times 10^{-3}$$

Where the values 0.7 Sv.Gy^{-1} is taken as the conversion coefficient from absorbed dose in air to effective dose received by adults and 0.2 is the outdoor occupancy factor.

3.3.3. Radium Equivalent Activity (R_{eq})

Due to a non-uniform distribution of natural radionuclides in the soil samples, the actual activity level of ^{226}Ra , ^{232}Th and ^{40}K in the samples can be evaluated by means of a common radiological index named the radium equivalent activity (R_{eq}). It is the most widely used index to assess the radiation hazards. This estimates that 370 Bq.kg^{-1} of ^{226}Ra , 259 Bq.kg^{-1} of ^{232}Th and 4810 Bq.kg^{-1} of ^{40}K produce the same gamma-ray dose rate. Therefore, R_{eq} can be obtained by the following relation [15]:

$$R_{\text{eq}} \text{ (Bq.kg}^{-1}\text{)} = A_{\text{Ra}} + 1.43A_{\text{Th}} + 0.077A_{\text{K}}$$

Where A_{Ra} , A_{Th} and A_{K} are the activity concentration of ^{226}Ra , ^{232}Th and ^{40}K in Bq/kg respectively. The permissible maximum value of the radium equivalent activity is 370 Bq.kg^{-1} which corresponds to an effective dose of 1 mSv for the general public.

3.3.4. External Hazard Index (H_{ex})

To limit the radiation exposure attributable to natural radionuclides in the samples to the permissible dose equivalent limit of 1 mSv.y^{-1} , the external hazard index based on that criterion have been introduced using a model proposed by Krieger (1981) which is given by [16,17].

$$H_{\text{ex}} = A_{\text{Ra}}/370 + A_{\text{Th}}/259 + A_{\text{K}}/4810$$

Where A_{Ra} , A_{Th} and A_{K} are the activity concentration of ^{226}Ra , ^{232}Th and ^{40}K in Bq.kg^{-1} , respectively. In order to keep the radiation hazard insignificant, the value of external hazard index must not exceed the limit of unity. The maximum value of H_{ex} equal to unity corresponds to the upper limit of radium equivalent activity 370 Bq.kg^{-1} [18].

3.3.5. The Representative level index (I_{yr})

This index can be used to estimate the level of γ -radiation hazard associated with the natural radionuclides. The

radiation level index may be defined as [19].

$$I_{\gamma r} = A_{Ra}/150 + A_{Th}/100 + A_K/1500$$

Where A_{Ra} , A_{Th} and A_K are the activity concentration of ^{226}Ra , ^{232}Th and ^{40}K in Bq.kg^{-1} , respectively. $I_{\gamma r}$ May be used to estimate the level of gamma radiation hazard associated with natural radionuclides in the investigated soil sample.

All the calculated health hazard parameters are gathered and displayed in Table (4).

Table 4. The radium equivalent (Ra_{eq}), the external hazard index (H_{ex}), the representative level index ($I_{\gamma r}$), the absorbed dose rate (D) in nGy/h and annual effective dose equivalent (AEDE) outdoor in $\mu\text{Sv.y}^{-1}$ for the studied samples.

Sample No.	Ra_{eq} (Bq/kg)	H_{ex}	$I_{\gamma r}$	Absorbed dose rate nGy.h^{-1}	AEDE (outdoor) $\mu\text{SV year}^{-1}$
Sh-1	121.92	0.33	0.86	55.74	68.36
Sh-2	151.34	0.41	1.06	69.16	84.82
Sh-3	201.00	0.54	1.41	91.34	112.02
Sh-4	173.74	0.47	1.23	79.60	97.62
Sh-5	210.97	0.57	1.49	96.49	118.34
Sh-6	243.13	0.66	1.71	110.89	135.99
Sh-7	215.30	0.58	1.52	98.52	120.82
Sh-8	207.42	0.56	1.47	95.15	116.69
Sh-9	287.46	0.78	2.02	131.36	161.10
Sh-10	243.38	0.66	1.71	111.18	136.36
Sh-11	455.35	1.23	3.15	205.60	252.15
Sh-12	261.68	0.71	1.84	119.44	146.49
Sh-13	257.17	0.69	1.78	116.05	142.32
Sh-14	193.70	0.52	1.36	88.31	108.30
Sh-15	265.00	0.72	1.86	120.71	148.04
Sh-16	273.01	0.74	1.91	124.26	152.39
Sh-17	362.83	0.98	2.54	165.30	202.72
Sh-18	517.64	1.40	3.57	233.45	286.30
Sh-19	350.25	0.95	2.42	158.34	194.19
Sh-20	261.05	0.71	1.82	118.63	145.49
average	262.67	0.71	1.84	119.48	146.53

As shown in Table (4), the estimated absorbed dose rates based on soil radioactivity are ranged from 55.74 to 233.45 nGy.h^{-1} with a mean value 119.48 nGy.h^{-1} . This mean value is higher than the worldwide mean value of 58 nGy/h [20].

The effective dose for the different locations varied from 68.36 to 286.3 $\mu\text{Sv.y}^{-1}$, with the mean value of 146.53 $\mu\text{Sv.y}^{-1}$, which is higher than the worldwide effective dose of 70 $\mu\text{Sv.y}^{-1}$ [21].

The values of Ra_{eq} are ranged from 121.92 to 517.64 Bq.kg^{-1} with an overall mean of 262.67 Bq.kg^{-1} . It can be seen that the Ra_{eq} values for all soil samples in the present work (with the exception of the values associated with samples no. Sh-11 and Sh-18) are lower than the accepted safety limit value of 370 Bq.kg^{-1} [22].

The calculated values of the external hazard index for all soil samples vary from 0.33 to 1.4 where the average value were found to be 0.71. The results show that the H_{ex} values for all soil samples (with the exception of the values associated with samples no. Sh-11 and Sh-18) are below the limit of unity. This result means that the radiation dose is

below the permissible limit of 1 mSv.y^{-1} recommended by ICRP [23]. It can be concluded that the radiological health risks to the people living in the studied area is significant at certain zones.

Furthermore, $I_{\gamma r}$ for most samples is greater than unity which indicates that Nugrus area represents to some extent hazardous area that needs special precautions in establishing any trials to extract or explore the existing important elements.

4. Conclusion

The present work is devoted to investigate the health hazards of the environmental rock and soil samples from Abu Rusheid area in south eastern desert of Egypt using γ -ray spectroscopy and XRF techniques. As a consequence, the activity concentration of ^{226}Ra , ^{232}Th , ^{40}K in the collected samples were evaluated. The absorbed dose rate, the radium equivalent, the external health hazard, the representative gamma index and the annual effective dose were also estimated and found to represent a significant radiological health hazard to people living in the investigated area. Therefore, safety rules and precautions should be applied for those workers in these fields. Also, the obtained results can serve as a radioactivity database for persons working in mining and milling activity in Abu –Rusheid area. Additionally, the samples were described by X-ray fluorescence and the obtained results prove that the majority of the samples contain considerable strategic elements concentrations such as Al, Fe, Mn and Zn that have considerable values in the advanced industrial domain.

Acknowledgement

Authors would like to cordially thank all the members of the nuclear physics group in the Physics Department, Faculty of Science, Ain-Shams University, Cairo, Egypt for their help.

References

- [1] N. K. Ahmed, A. M. El Arabi, H. M. Mahmoud, and K. Salabel-din, "Measurement of natural radioactivity and its significant hazards of some hematite samples in Eastern Desert, Egypt," *Build. Environ.*, vol. 42, no. 6, pp. 2263–2267, 2007.
- [2] A. Sroor, S. M. El-Bahi, F. Ahmed, and A. S. Abdel-Haleem, "Natural radioactivity and radon exhalation rate of soil in southern Egypt," *Appl. Radiat. Isot.*, vol. 55, no. 6, pp. 873–879, 2001.
- [3] P. Chiozzi, V. Pasquale, and M. Verdoya, "Naturally occurring radioactivity at the Alps-Apennines transition," *Radiat. Meas.*, vol. 35, no. 2, pp. 147–154, 2002.
- [4] T. E. Myrick, B. A. Berven, and F. F. Haywood, "Determination of Concentrations of Selected Radionuclides in Surface Soil in the US.," *Health Phys.*, vol. 45, no. 3, pp. 631–642, 1983.

- [5] A. A. Abdel-Razek, Y. A.; Bakhit, A F;Nada, "Measurements of the Natural Radioactivity along Wadi Nugrus, Egypt," *Radiat. Phys. Prot. Conf.*, no. November, pp. 15–19, 2008.
- [6] F. Adams, K. Janssens, and A. Snigirev, "Microscopic X-ray fluorescence analysis and related methods with laboratory and synchrotron radiation sources," *J. Anal. At. Spectrom.*, vol. 13, no. 5, pp. 319–331, 1998.
- [7] O. A. A. Eletta, "Determination of some trace metal levels in Asa river using AAS and XRF techniques," *Int. J. Phys. Sci.*, vol. 2, no. 3, pp. 56–60, 2007.
- [8] N. Ibrahim, "Natural activities of ^{238}U , ^{232}Th and ^{40}K in building materials," *J. Environ. Radioact.*, vol. 43, no. 3, pp. 255–258, 1999.
- [9] Unsear report, "Sources and Effects of Ionizing radiation," 1993.
- [10] I. F. Al-Hamarneh and M. I. Awadallah, "Soil radioactivity levels and radiation hazard assessment in the highlands of northern Jordan," *Radiat. Meas.*, vol. 44, no. 1, pp. 102–110, 2009.
- [11] M. Tzortzis, "Determination of thorium, uranium and potassium elemental concentrations in surface soils in Cyprus," *J. Environ. Radioact. J. Environ. Radioact.*, vol. 77, no. 3, pp. 325–338, 2004.
- [12] M. Bashir, I. Ibeanu, Y. Zakari, and U. Sadiq, "Assessment of Radiological Risk in Flooded Soil Samples of Kudenda, Kaduna State Nigeria," *Int. J. Eng. Sci. Inven.*, pp. 69–74, 2013.
- [13] Unsear report, "Exposures from natural radiation sources," 2000.
- [14] N. Damla, U. Cevik, A. I. Kobya, A. Celik, N. Celik, and R. Van Grieken, "Radiation dose estimation and mass attenuation coefficients of cement samples used in Turkey," *J. Hazard. Mater.*, vol. 176, no. 1–3, pp. 644–649, 2010.
- [15] H. M. Diab, S. A. Nouh, A. Hamdy, and S. A. El-Fiki, "Evaluation of natural radioactivity in a cultivated area around a fertilizer factory," *J Nucl Radiat Phys*, vol. 3, no. 1, pp. 53–62, 2008.
- [16] M. Tufail, T. Hamid, and others, "Natural radioactivity hazards of building bricks fabricated from saline soil of two districts of Pakistan," *J. Radiol. Prot.*, vol. 27, no. 4, p. 481, 2007.
- [17] R. Krieger, "Radioactivity of construction materials," *Betonw. Fert. Techn.*, vol. 47, p. 468, 1981.
- [18] S. Turhan and L. Gunduz, "Determination of specific activity of ^{226}Ra , ^{232}Th and ^{40}K for assessment of radiation hazards from Turkish pumice samples," *J. Environ. Radioact. J. Environ. Radioact.*, vol. 99, no. 2, pp. 332–342, 2008.
- [19] A. A. Kinsara, E. I. Shabana, and M. T. Qutub, "Natural radioactivity in some building materials originating from a high background radiation area," *Int. J. Innov. Educ. Res.*, vol. 2, no. 6, pp. 70–78, 2014.
- [20] European Commission, "Radiological Protection Principles concerning the Natural Radioactivity of Building Materials," 1999.
- [21] Y. Örgün, N. Altinsoy, S. Y. Sahin, Y. Güngör, A. H. Gültekin, G. Karahan, and Z. Karacik, "Natural and anthropogenic radionuclides in rocks and beach sands from Ezine region (Canakkale), Western Anatolia, Turkey," *Appl. Radiat. Isot.*, vol. 65, no. 6, pp. 739–747, 2007.
- [22] M. Belivermis, Onder Kihc, Y. Çotuk, and S. Topcuoglu, "The effects of physicochemical properties on gamma emitting natural radionuclide levels in the soil profile of Istanbul," *Environ. Monit. Assess.*, vol. 163, no. 1–4, pp. 15–26, 2010.
- [23] U. N. S. C. on the Effects of Atomic Radiation, *Sources and effects of ionizing radiation: sources*, vol. 1. United Nations Publications, 2000.